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Biodegradation and detoxification of low-density polyethylene by an indigenous strain *Bacillus licheniformis* SARR1

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ABSTRACT

Plastics are synthetic polymers, which are frequently used in daily life for a wide range of purposes. These plastic wastes are accumulated and generate plastic pollution in the environment. It needs many years for complete deterioration in the environmental conditions. Biodegradation is the most promising method to treat plastic pollution, as the microorganisms utilize the low-density polyethylene (LDPE) as a sole source of carbon, and this indicates an innovative approach to manage the problem of plastic waste. The isolate SARR1 was identified as *Bacillus licheniformis* using the National Center for Biotechnology Information (NCBI) database. The isolate SARR1 showed an LDPE removal rate (*K*) of 0.069 g day⁻¹ with a half-life of approximately 335.32 days to degrade LDPE strips. The biomass production was $0.98 \pm 0.006 \text{ gl}^{-1} (X_m)$ during the incubation of 30 days, and the percentage of crystallinity was significantly decreased from 71.69% to 50.78% due to biodegradation. The esterase and lipase activity of isolate SARR1 was studied using UV visible spectroscopy. The gas chromatography-mass spectrometry analysis confirmed the synthesis of acetone, diazene dimethyl, and carbamimidothioic acid, 1-methylethyl ester with different peak area percentages of 23.38%, 65.58%, and 11.04%, respectively. Seed germination study showed that the compounds formed after biodegradation of LDPE by bacterial strain SARR1 were eco-friendly.

1. INTRODUCTION

Plastic polymers are a large family that belongs to the monomer unit of ethylene $(-CH_2-CH_2-)_n$ and are used in packaging, electronics, and textiles [1]. Therefore, these are recalcitrants that accumulate in the environment and pollute the sediments immensely around the earth. Due to plastic pollution, a sharp public concern has been seen in a massive quantity that poses an ecological threat such as leaching, fragmentation, additive migration due to quickly colonization by the diversity of microorganisms, and termed as plastisphere. Furthermore, the biodegradation of highly durable polymers such as low-density polyethylene (LDPE) is less suitable for microorganisms due to its higher redox potential, requiring more energy to break the C–C bonds. The worldwide

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Anita Rani Santal, Department of Microbiology, Maharshi Dayanand University, Rohtak, India. E-mail: anita.micro @ mdurohtak.ac.in utilization of polyethylene rapidly increases by over 320 million tons every year and seems to be doubled by 2034 [2]. The pollution caused by polyethylene includes blocking drains, water, soil, air contamination, sewage lines around the urban areas due to non-degradability and is not beneficial to the topsoil [3].

LDPE is usually unreactive at room temperature, and with the addition of oxidizing agents, they become reactive. In most cases, plastics decompose aerobically, which produces carbon dioxide and water as a product. However, anaerobically, it produces methane, water, and carbon dioxide. The outcomes of oxidative degradation of polyethylene were studied in a controlled condition and found to consist of harmful chemicals that are dangerous for humans and the environment [4]. Therefore, the bioremediation method boosts the degradation rate without damaging the environment. The bioremediation process removes recalcitrants from their origin and eliminates the ecosystem's pollutants [5]. This process is influenced by many factors such as the number of microorganisms that participated, contact time, nutrient availability, temperature, pH, consortium formulation, and pollutant

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characteristics. Various researchers explored polyethylene degrading bacteria such as Bacillus spp., Pseudomonas spp., Acinetobacter sp., Brevibacillus sp., Flavobacterium spp., Ralstonia spp., Micrococcus spp., Microbacterium sp., and Nocardia sp. [6]. Enzymes like dehydrogenases and esterases, the AlkB family of hydrolases, initiate the LDPE oxidation process [7]. The laccase enzyme (copper-binding bacterial enzyme) can also degrade LDPE. This enzyme reduces the LDPE polymer's molecular weight (40,000 Daltons) with a simultaneous increase in a keto-carbonyl index [8]. A well-known example, Pseudomonas aeruginosa E7, releases alkane hydroxylases, which showed 30% of LDPE degradation [9]. The depolymerases enzymes, which are extracellular, break the complex synthetic polymers into monomers and dimers [10]. Thus, the microbes utilize these short-structured oligomers as carbon sources [11]. The PETase enzyme secreted by Ideonella sakaiensis hydrolyzes Polyethylene terephthalate (PET) into mono hydroxyethyl terephthalate, terephthalate, and ethylene glycol, which are further metabolized by bacteria through the pathway of β -oxidation [12]. Bacillus and Brevibacillus sp. secrete enzyme protease, which catalyzes the proteolysis of LDPE [13]. The bacteria Streptomyces viridosporus T7A, Streptomyces badius 252, and Streptomyces setonii 75Vi2 secretes enzymes like monooxygenase, dioxygenase dehydrogenase, which break down complex polymer LDPE [14]. Various techniques such as X-ray diffraction (XRD), scanning electron microscopy (SEM), and Fourier-transform infrared spectroscopy (FTIR) analysis are often used to detect LDPE structure changes [15]. In the present study, the ability to degrade LDPE has been investigated by a bacterium SARR1 isolated from waste disposal sites of Haryana, India. The LDPE degradation was studied using SEM, XRD, FTIR, and gas chromatography-mass spectrometry (GC-MS).

2. MATERIALS AND METHODS

2.1. Sample Collection and LDPE Used

In the current study, LDPE was procured from the local market, Rohtak, Haryana. LDPE thickness was 10 μ , and thin strips were cut into 3 × 3 cm. These LDPE strips were disinfected and washed with 75% ethanol and air-dried. It was used as the sole carbon source in the media for the growth of bacteria. The bacterial culture was isolated from the soil sample collected from the waste disposal site, Ismailabad, Pehowa (Haryana), India. The soils samples were then processed within 24 hours of collection and stored at 4°C.

2.2. Isolation and Cultivation of Bacterial Isolates

The enrichment culture technique was implemented to isolate bacteria that can degrade LDPE. In this technique, 10 g of soil sample was suspended in 90 ml of mineral salt broth media, amended with LDPE strips, and incubated for 30 days at 37°C under static conditions, respectively. After incubation, 100 μ l inoculums were spread on agar plates and purified using the streak plate method [16].

2.3. Molecular Characterization and Bacterial Identification

The bacterial strain SARR1 was identified using 16S rRNA gene sequencing. The 16S rRNA gene was amplified using universal primers, i.e., fD1 (5'-AGAGTTTGATCCTGGCTCAG-3') and rP2 (5'-ACGGCTACCTTGTTACGACTT-3') [17]. The

universal primers were complementary to the conserved regions at the beginning of the gene sequence and either of the whole sequence (1,550 bp). Therefore, the sequence of the variable region was used as comparative taxonomy. The polymerase chain reaction (PCR) composition of 50 μ l total volume was 50 ng genomic DNA, primer fD1 and rP2 each (20 pmol), Deoxynucleotide triphosphates (dNTPs) (Sigma) (200 M each), Taq polymerase buffer (10×), and Taq DNA polymerase (Sigma) (2.5 U) and PCR grade water. The PCR reaction was carried out using Thermal Cycler (Applied biosystems model 2720), and the amplification was done for 35 cycles [18]. The PCR amplified product was resolved on the 1.5% agarose gel electrophoresis and then purified using a genetix nucleo-pore PCR clean-up kit as per the manufacturer's instructions.

The nucleotide sequence of 16S rDNA was done by Biokart India Pvt. Ltd., Bangalore, Karnataka, India. The obtained nucleotide sequences were then analyzed on the NCBI Basic Local Alignment Search Tool (BLAST) tool. A phylogenetic tree of the obtained sequence and the retrieved 16S rRNA sequences was constructed using the neighbor-joining method [19]. MEGA X software was used for the evolutionary analysis [20,21].

2.4. Biodegradation Assay of LDPE and Kinetic Model

The growth kinetics of SARR1 was studied at a regular time interval both with and without LDPE strips. The MS broth with and without LDPE strips was inoculated with SARR1 and incubated at 37°C for 30 days, and Optical Density (OD) was measured at 600 nm [22]. The weight-loss method was used to determine LDPE degradation's rate constant by the first-order kinetics model [23]. The rate constant for the removal of LDPE was followed through half-life (t//2) [24]. The biomass growth kinetics of well-grown isolate SARR1 was studied using the Logistic model equation.

2.5. Enzyme Activity

The cell-free supernatant was used to demonstrate the enzyme activities of Esterase and lipase by UV visible spectrophotometer. The samples were extracted and centrifuged at $3,427 \times g$ at 30° C to derive cell-free supernatant from biodegradation assay at 10, 20, and 30 days [25].

2.5.1. Esterase and lipase enzyme assay

To check the activity of esterase and lipase enzyme, the substrate *p*-nitrophenyl butyrate and palmitate were used respectively, added in 1 ml acetonitrile, and made up the volume of 100 ml with addition 0.1 M phosphate buffer solution (pH 7.0) [25]. The reaction mixture was prepared by adding 0.8 ml phosphate buffer solution (0.1 M, pH 7.0), 0.066 ml supernatant devoid of cells, and the substrate of 0.134 ml incubated at 30°C for 30 minutes. The sample changed its color and was measured with respect to control as a blank using a UV visible spectrophotometer at 410 nm [26].

2.6. LDPE Degradation Analysis

The evaluation of biodegradation and structural changes of LDPE by the bacterial isolate SARR1 was analyzed using SEM, FTIR, and XRD.

2.6.1 LDPE degradation analysis using SEM

After incubation of 30 days, the treated and untreated LDPE strips were rinsed with 2.0% (v/v) Sodium dodecyl sulphate (SDS), followed by distilled water and 70% (v/v) ethanol. The LDPE strip was coated with gold for the 40 seconds (Ion sputter JFC-1100) and analyzed under a high-resolution SEM (JSM-6100SEM) [27].

2.6.2. LDPE degradation analysis using FTIR

The treated and untreated LDPE strip was rinsed with 70% ethanol and air-dried overnight [28]. The samples were analyzed under the attenuated total reflectance-FTIR spectrophotometer (Model Alpha, Bruker, Germany) for the chemical and structural modifications induced by isolate SARR1 on the LDPE strip scanned from the scan 4,000 to 500 cm⁻¹ at room temperature.

2.6.3. LDPE degradation analysis using XRD

The XRD analysis of LDPE was done using an X-ray diffractometer and operated using Cu (K α) radiation ($\lambda = 1.5406$ Å). The radiation scattered and registered in the angular interval (2 θ) was from 0° to 80°. A current of 15 mA and a voltage of 30 kV were used (MiniFlex 2 goniometer). The XRD diffraction peaks were investigated under constant operating conditions on the XRD diffractometer, and the data were recorded. The crystallinity percentage was calculated using the following formula [29].

% Crystallinity
$$\frac{I_c}{I_c + I_a} \times 100$$

 I_c is the total area of crystalline peaks, and $I_c + I_a$ is the total area of all peaks.

2.6.4. LDPE degradation analysis using GC-MS

The degraded LDPE products were analyzed by GC-MS using a Thermo Trace 1300GC coupled with Thermo TSQ 8000 Triple Quadrupole MS, equipped with BP 5MS (30 m \times 0.25 mm, 0.25 µm) column. Helium gas as a carrier and di-ethyl ether as a solvent were used [30]. The LDPE strips were extracted from the inoculated broth culture. The cell culture was centrifuged at 6,093 \times g for 5 minutes at room temperature, and the cell debris was removed. The supernatant was filtered by Whatman filter paper, and LDPE products were collected using diethyl ether as a solvent. 10 ml filtrate was then dissolved in an equal volume of diethyl ether and separated using a separating funnel [30]. Finally, 1.0 ml of dissolved LDPE degraded products were injected in GC-MS. The metabolites were identified based on mass spectra and National Institute of Standards and Technology (NIST) library 2.0 [31].

2.7. Phytotoxicity Analysis of LDPE Degraded Products

Seed germination bioassay is the most common method used to evaluate the phytotoxicity of any treated samples [32]. The phytotoxicity tests were performed to assess the toxicity of the treated LDPE sample. The experiment was performed using sterilized Petri plates lined with circular pieces of Whatman no.1 filter paper. The 1.0 ml of the treated and untreated LDPE sample was poured into the Petri plates, respectively. The seeds of *Vigna radiata* were disinfected with 70% ethanol, soaked for 30 minutes in respective LDPE samples. Afterward, seeds were placed on the moistened filter papers at room temperature for 48 hours [33]. The phytotoxicity of seeds was estimated by comparing the germination of seeds in both treated and untreated samples, respectively [34].

3. RESULTS AND DISCUSSION

3.1. Isolation of LDPE Degrading Bacteria

Bacterial cultures were isolated and screened out based on the weight loss method. An isolate *Bacillus licheniformis* SARR1 showed $33.32\% \pm 1.39\%$ biodegradation efficiency at 37° C and pH 7.0. The isolate SARR1 was more capable of consuming and assimilating the LDPE strips as a sole carbon source and thus showed biodegradation in aerobic conditions. However, after utilization, the conversion of the polymeric compound into simple oligomers occurred, and byproducts such as CO₂ and H₂O also might be released. Researchers have earlier reported various isolates, but in contrast, to present isolate SARR1, the capability of degradation was less in prior reported cultures. *Enterobacter cloacae* AKS7 sp. was the effective strain for LDPE degradation from an agricultural site inoculated in mineral salt broth medium amended with LDPE strips as a sole carbon source [35].

3.2. 16S rRNA Gene Sequence and Phylogenetic Analysis

3.2.1. 16S rRNA gene sequence analysis

For the identification of potent bacterial isolate SARR1, the 16S rRNA gene sequencing method was used. The amplification of the PCR product of the 16S rRNA gene of the isolate SARR1 was approximately 1,210 bp (Fig. 1). The obtained 16S rRNA nucleotides were then subjected to BLAST analysis for the identification of an isolate. The partial sequence of the 16S rRNA gene was submitted to NCBI with the accession number MT974150. The 16S rRNA gene sequences of isolate SARR1 and the retrieved sequences from the NCBI database and Remote Desktop Pro (RDP) database showed sequence homology to B. licheniformis strain DSM 13 and strain BCRC 11702 with a sequence similarity of 99.92%. The phylogenetic tree of the isolate SARR1 was constructed using the neighbor-joining method with the closest neighbor strain (Bacillus piscis strain 16MFT21 and Bacillus paralicheniformis strain KJ-16), which showed the isolate belonged to B. licheniformis (Figs. 1 and S1). Similarly, LDPE degrading bacterial isolates had earlier been isolated and identified as Pseudomonas citronellolis EMBS027 [36], Microbulbifer hydrolyticus IRE-31 [37], and P. aeruginosa ISJ14 [6].

3.3. Biodegradation Assay of Isolate and Its Growth Kinetic Model

The biodegradation assay of LDPE was studied for 30 days, and percent weight loss was recorded for growth kinetic analysis at a time interval of 5 days of incubation (Fig. 2). The bacterial growth was increased in the MS media amended with LDPE compared with the presence of limited glucose concentration (0.5 g/l) in



Figure 1: The phylogenetic relationship of isolate SARR1 with the retrieved 16S rRNA sequences from the NCBI BLAST was constructed using the neighbor-joining method. The evolutionary analysis was conducted in MEGA X.

MS broth media. The altered growth rate was observed in the limited glucose concentration, which its metabolite pool might control compared to environmental conditions. In this way, the strain SARR1 showed the potential to grow and consume LDPE as a sole carbon source. The maximum degradation of LDPE was then evaluated as $33.32\% \pm 1.39\%$ under optimized conditions. Similarly, the isolate P. aeruginosa ISJ14 degraded only 6.5% LDPE within 2 months of incubation [38] while Brevibacillus borstelensis 707 degraded 2.5% after 1 month of incubation [39]. The Streptomyces sp. growth was observed in the presence of LDPE as a carbon source showed the potential to degrade LDPE [22]. The isolate SARR1 had an LDPE removal rate (K) of 0.069 g day⁻¹ through a longer half-life $(t^{1/2})$ approximately required 335.32 days to degrade LDPE strips into its half in contrast to P. aeruginosa ISJ14 that showed a removal rate of LDPE of 0.0012 g day⁻¹ and degrade into its half in 577.5 days [38]. Similarly, Bacillus sp. and Rhodococcus sp. showed the removal rate of microplastic 0.0019 g day⁻¹ with a half-life of 363 days [24].

The biomass production of strain SARR1 in the LDPE enriched medium was $0.98 \pm 0.006 \text{ gl}^{-1}(X_w)$ with a weight loss percentage of

32.15% \pm 2.27% after incubation of 30 days. Montazer *et al.* [40] reported biomass production of 0.15 \pm 0.007 gl⁻¹ for *Micrococcus luteus* IRN20 and 0.37 \pm 0.05 gl⁻¹ of biomass production for *Cupriavidus necator* H16 within 21 days on LDPE containing media.

3.4. Enzymatic Assay During Degradation of LDPE

The production of esterase and lipase enzymes was determined and revealed during the degradation of LDPE by an isolate SARR1. The isolate SARR1 showed the production and activity of esterase and lipase enzymes on different days (Fig. 3a and b). The earlier studies demonstrated that lipase and esterase enzymes were the lipolytic enzyme that belonged to the hydrolases class of enzymes that hydrolyzed fatty acid and glycerol. They also participate in the catabolic activities of hydrocarbon degradation [25]. The action of enzymes connected with respiration and energy production pathways demonstrates the metabolic state of bacteria [26]. Kumar *et al.* [41] identified that *Bacillus* sp. was the best producer of lipase and esterase enzyme for LDPE degradation. Rani et al.: Biodegradation and detoxification of low-density polyethylene by an indigenous strain Bacillus licheniformis SARR1 2022;10(01):9-21



Figure 2: (a) Biodegradation assay demonstrating weight loss percentage of LDPE at regular intervals of time (b) Growth kinetics of the isolate SARR1 with LDPE and with limited glucose concentration (0.5 g/l).



Figure 3: Enzyme activity at different days was optimized by UV visible spectroscopy (a) Esterase activity (b) Lipase activity.

3.5. Evaluation of Structural Changes of LDPE

3.5.1. LDPE analysis using SEM

The SEM analysis of the surfaces of treated and untreated LDPE strips was done at different resolutions $(500 \times \text{ and } 2,500 \times)$ after 30 days of incubation. The surface erosion cracks, folds, and big holes were seen on the surface of LDPE strips (Fig. 4). Hence, the disintegration of the smooth LDPE surface was observed and recorded. Several researchers have well studied the microbial degradation mechanism of LDPE. The bacteria utilized the oxidized region and solubilized it, leading to pits formation on the surface of LDPE [42,43]. It was depicted that due to enzymatic activity, oxidation of LDPE. Mukherjee *et al.* [44] also observed

a disintegrated surface of LDPE after treatment with *Pseudomonas fluorescens*. Microbial attachment on the LDPE surface indicates its strong adherence and LDPE utilization capacities. The cracks and pits on the LDPE surface after treatment with *Microbulbifer hydrolyticus* IRE-31 and *Streptomyces albogriseolus* LBX-2 confirmed LDPE degradation [37].

3.5.2. LDPE analysis using FTIR

As the LDPE consists of ethylene groups, the major infrared (IR) band has $-CH_2$ - stretching and bending vibrations. The change in the peak values and functional groups supported the conformational change of the polymer surface. The IR spectra of the treated LDPE surfaces showed peaks at 2,845.08, 1,744.36-



Figure 4: SEM analysis of LDPE biodegradation at different resolutions (500×, 2,500×) (a) control sample (b) Different resolutions of biodegraded LDPE treated with isolate SARR1 revealing surface changes.

1,463.13, and 718.84-626.59 cm⁻¹ for C-H deformation vibration CH₂ scissoring vibration, -CHO stretching, and strong alkyl structure, respectively. The strong bond pattern was observed at 2,913.98, which showed C-H stretching vibration asymmetrically (Fig. 5). The new peaks with a marked shift at 1,500-1,540 cm⁻¹ (C=C stretching vibration, O-H stretching: carboxylic acid) and 1,000–1,100 cm⁻¹ (-CHO stretching, C=O stretching) was observed in the IR spectra of LDPE degradation. The disappearance of the peaks in the range of 2,000 to 2,500 cm⁻¹ was also observed and depicted that the N-H and C-H vibrations represented the amide group addition. This proved that bacteria are capable of utilizing long, hydrophobic polyethylene chains. The formation of new peaks at 2,236.00-2,269.00 cm⁻¹ (O-H stretching: carboxylic acid), 1,555.44 cm⁻¹ (C=C stretching vibration), 788.01-788.86 cm⁻¹ (C-H rocking vibration: CH₂) on IR spectra of degraded LDPE were also observed by Kunlere et al. [43]. Similarly, the peak at 2,660 cm⁻¹ (-CHO stretch) was disappeared in *Bacillus* amyloliquefaciens treated LDPE, while the new peak was demonstrated at 939 cm⁻¹ (O-H bend) which showed LDPE depolymerization [45].

3.5.3. LDPE analysis using XRD

The XRD analysis showed that the distinguished highest peak in the control and treated sample chromatogram was observed at 21.7 of the angular position 2θ (Fig. 6a and b). The new peaks were observed in the range of 20–30 of angular position 2 θ . After degradation, the minor shift in the peak's position at 20.7, 20.92, 21.8, and 22.62 was seen. The crystallinity percentage was decreased from 71.69% (control) to 50.78% in the treated sample within 30 days of incubation (Fig. 6c). The amorphous region of LDPE was found disintegrated in the treated sample. *Bacillus siamensis* and *Bacillus wiedmannii* decreased the crystallinity percentage of LDPE after 90 days of incubation [46]. The crystallinity of LDPE was decreased by treating with the mixed cultures of *Aspergillus* sp. and *Lysinibacillus* sp. [29], *Vibrio* sp., and *Aspergillus* sp. [16].

3.5.4. LDPE analysis using GC-MS

The GC-MS analysis showed that compounds like acetone, diazene, dimethyl, and carbamimidothioic acid, 1-methyl ethyl ester were obtained with peak area percentages in the treated sample 23.38%, 65.58%, and 11.04%, respectively (Table S1, Fig. S2). A new peak of Carbamimidothioic acid, 1-methylethyl ester appeared in the treated LDPE while the compounds like acetone, 1,2,4,5-tetroxane, 3,3,6,6-tetramethyl, 1,4-benzenedicarboxylic acid. bis(2ethylhexyl)ester, 1,3-benzenedicarboxylic acid. bis(2ethylhexyl) ester, isophthalic acid, 2-ethylhexyl octyl ester, terephthalic acid, 4 octyl ester, terephthalic acid, and 2-Ethylhexyl octyl ester were obtained with peak area percentages of 3.11%, 6.11%, 1.05%, 1.13%, 1.25%, 1.73%, 2.58%, 2.12%, 25.58%, and 55.35%, respectively in the control sample (Table S1, Fig. S3). The compound Carbamimidothioic acid, 1-methylethyl ester, can act as a potent anti-cancer agent [47]. After further degradation of carboxylic acid by an isolate, SARR1 might produce alkane

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Figure 5: FTIR spectra of LDPE biodegradation after 30 days of the incubation period (a) Control sample (b) LDPE sample treated with isolate SARR1.

compounds such as butane, 1,2,4,5-tetroxane, and methane, shown in the chromatogram of the treated sample. The production of such compounds was confirmed by observing the IR spectra obtained after the degradation of LDPE. Similarly, new peaks with retention times 7.46 and 17.28 of 2-butene, 2-methyl, and ethane were earlier observed in *Acinetobacter baumannii* treated LDPE [48]. However, *P. aeruginosa* treated LDPE film showed a long chain of esters, hydrocarbons, aldehydes, oxygenated chemical compounds by gas chromatography [30]. On the other hand, the *Lysinibacillus fusiformis* strain VASB14/WL treated LDPE film produced 1- trimethylsilyl methanol, hexadecanoic acid, and 1,2,3,4- tetra methylbenzene, whereas *Bacillus cereus* strain VASB1/TS resulted in the formation of 1-trimethylsilyl methanol, 1,2,3-trimethyl benzene, and hexadecanoic acid [49].

3.6. Phytotoxicity Analysis of LDPE Degraded Products

The primary aim was to assess the LDPE degradation's phytotoxicity and metabolites after degradation by bacterial isolate SARR1. The phytotoxicity analysis showed that the seed germination in *V. radiata* was high (100% \pm 0%) when irrigated with SARR1 treated LDPE samples compared to untreated (44.49% \pm 2.47%). Similarly, the germination speed was also quite high in the treated LDPE samples (Fig. 7). So, it indicated that the phytotoxicity of the treated samples was reduced and showed no inhibition on seed germination, and hence became eco-friendly. Therefore, the isolate *B. licheniformis* SARR1 showed a positive effect on the germination of *V. radiata*. Pathak and Kumar [50] also reported a positive effect on the seed germination, shoot height, and root length of *Vigna mungo* when irrigated with *Bacillus* sp. V8 and



Figure 6: XRD spectra of the LDPE films after incubation of 30 days (a) control (b) treated with isolate SARR1 (c) % Crystallinity.

Pseudomonas sp. C25 treated LDPE samples. The phytotoxicity standard tests have also been performed to determine the toxicity of polyethylene on *Lepidium sativum*, which resulted in a potential threat to the growth of plants [51]. The pollutant causes oxidative stress, and as a result, plants produce a high concentration of reactive oxygen species in their tissues and ultimately leads to cell damage [52].

4. CONCLUSION

Plastic waste and polyethylene are generally major ecological threats to the environment due to their non-biodegradable behavior. Therefore, significant attention has been placed on the biodegradation of polymers and the identifications of microorganisms with great potential. The current investigations demonstrated that the bacterium SARR1 develops substantially enhanced biodegradation over the surface of LDPE by exploiting the hydrophobicity of the cell surface. The biodegradation efficiency (%) of the strain SARR1 was $33.32\% \pm 1.39\%$ at temp. 37° C and pH 7.0. The biomass production by isolate SARR1



Figure 7: Phytotoxicity analysis of LDPE on seed germination of *V. radiata* (a) control (b) treated sample (c) germinated speed and % germination of seeds.

in the LDPE enriched medium was $0.98 \pm 0.006 \text{ gl}^{-1} (X_m)$ after incubation of 30 days, and the LDPE removal rate (*K*) was 0.069 g day⁻¹ with approximately 335.32 days of a half-life to degrade LDPE strips.

Moreover, the GC-MS analysis showed a new compound, carbamimidothioic acid, 1-methylethyl ester, formed during the degradation process, which has been reported as anticancer property. The SEM analysis of treated and untreated LDPE samples showed clear degradability of LDPE. Enzymatic study reveals that the esterase and lipase enzymes are supposed to be a part of LDPE degradation. The degraded products released from SARR1 treated LDPE showed no phytotoxic effect and are hence eco-friendly. This study would enable to develop more efficient microbial consortium having LDPE degradation capability. Moreover, to understand the mechanism of LDPE biodegradation, investigations towards the metabolic pathways and their enzymatic reactions are needed.

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AUTHOR CONTRIBUTIONS

All authors made substantial contributions to conception and design, acquisition of data, or analysis and interpretation of data; took part in drafting the article or revising it critically for important intellectual content; agreed to submit to the current journal; gave final approval of the version to be published; and agree to be accountable for all aspects of the work. All the authors are eligible to be an author as per the international committee of medical journal editors (ICMJE) requirements/guidelines.

CONFLICTS OF INTEREST

The authors report no financial or any other conflicts of interest in this work.

ETHICAL APPROVALS

This study does not involve experiments on animals or human subjects.

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SUPPLEMENTARY MATERIAL

Tuble S111 buk report of Ge Mis in control sumple und in control D51 D with S111 buk report of Ge Mis in control sumple					
S. No.	RT	Peak area	Area %	Peak height	Compound
1	3.18	19,406,540.7	3.11	4,437,044	Acetone
2	3.54	38,117,089.81	6.11	5,524,002	1,2,4,5-Tetroxane, 3,3,6,6-tetramethyl
3	3.8	6,535,000.36	1.05	1,028,322	Methane, oxybis[dichloro
4	19.47	7,034,617.18	1.13	948,828	1,4-benzenedicarboxylic acid, bis(2-ethylhexyl)ester
5	19.77	7,763,375.57	1.25	1,122,498	1,3-benzenedicarboxylic acid,bis(2-ethylhexyl) ester
6	20.18	10,796,187.89	1.73	1,977,727	Isophthalic acid, 2-ethylhexyl octyl ester
7	20.32	16,069,206.71	2.58	2,175,560	Terephthalic acid, di(4octyl) ester
8	20.53	13,200,532.32	2.12	1,735,733	Terephthalic acid, 4-octyl octyl ester
9	20.99	159,428,786	25.58	1E + 07	Terephthalic acid, 2-ethylhexyl octyl ester
10	21.44	34,502,0571.4	55.35	1.9E + 07	1,3-Benzenedicarboxylic acid,bis(2ethylhexyl) ester
Peak report of GC-MS treated LDPE with SARR1					
1	3.07	11,423,334.14	23.38	1,233,302	Acetone
2	3.39	32,033,695.3	65.58	3,791,467	Diazene, dimethyl
3	3.9	5,393,139.71	11.04	510,336	Carbamimidothioic acid, 1-methylethyl ester





FN597644.1:161663..163212 Bacillus amyloliquefaciens (T) DSM7 FN597644.1:167573..169122 Bacillus amyloliquefaciens (T) DSM7 FN597644.1:97406..98955 Bacillus amyloliquefaciens (T) DSM7 FN597644.1:91519..93068 Bacillus amyloliquefaciens (T) DSM7 FN597644.1:91526..93063 Bacillus amyloliquefaciens (T) DSM7 FN597644.1:97413..98950 Bacillus amyloliquefaciens (T) DSM7 FN597644.1:161670..163207 Bacillus amyloliquefaciens (T) DSM7 FN597644.1:167580..169117 Bacillus amyloliquefaciens (T) DSM7 FN597644.1:172935..174472 Bacillus amyloliquefaciens (T) DSM7 FN597644.1:918302..919839 Bacillus amyloliquefaciens (T) DSM7 FN597644.1:complement(2920237..2921774) Bacillus amyloliquefaciens (T) DSM7 CP000002.3:complement(3121757..3123307) Bacillus licheniformis (T) ATCC 14580 DSM 13 CP000002.3:920927..922477 Bacillus licheniformis (T) ATCC 14580 DSM 13 CP000002.3:158298..159848 Bacillus licheniformis (T) ATCC 14580 DSM 13 MT974150.1 Bacillus licheniformis strain SARR1 CP000002.3:95342..96892 Bacillus licheniformis (T) ATCC 14580 DSM 13 CP000002.3:34600..36149 Bacillus licheniformis (T) ATCC 14580 DSM 13 CP000002.3:9902..11451 Bacillus licheniformis (T) ATCC 14580 DSM 13 CP000002.3:611992..613541 Bacillus licheniformis (T) ATCC 14580 DSM 13 AF302118.1:1..1410 Bacillus sonorensis (T) NRRL B-23154 AJ831843.1:1..1494 Bacillus aerius (T) type strain:24K AB021192.1:1..1513 Bacillus mycoides (T) ATCC6462

Figure S1: The phylogenetic relationship of isolate SARR1 with the retrieved 16S rRNA sequences from the RDP database was constructed using the neighbor-joining method. Evolutionary analyses were conducted in MEGA X.



b

Figure S2: GC-MS analysis of LDPE film (a) LDPE film before treatment with the isolate SARR1 as control (b) LDPE film treated with isolate SARR1.

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Figure S3: GC-MS chromatogram of LDPE-degraded compounds synthesized during treatment with isolate SARR1 (a) Acetone (b) Diazene, dimethyl (c) Carbamimidothioic acid, 1-methylethyl ester.